

## Reef fish community structure on three islands of Itaipu, Southeast Brazil

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We provide here the first assessment of the composition, abundance and distribution of rocky reef fishes of Itaipu Sound, Rio de Janeiro, off the southeastern Brazilian coast. Monthly visual censuses, benthic quadrats and chain link transects were conducted over one year on rocky reefs at three islands (Menina, Mãe and Pai). A total of 2466 individual fish, belonging to 29 families and 42 species were recorded. The most abundant fish species were *Parablennius pilicornis*, *Haemulon steindachneri*, *Orthopristis ruber* and *Diplodus argenteus*. Sheltered and complex habitats showed the most abundant and diverse fish populations. There was a major significant separation between sampling sites and a secondary seasonal pattern. The three sites showed similar fish communities with locally structured environmental gradients according to their inherent habitat characteristics.

Apresentamos aqui o primeiro estudo da composição, abundância e distribuição de peixes recifais nos recifes rochosos da enseada de Itaipu, Rio de Janeiro, Brasil. Foram realizados mensalmente censos visuais, quadrats de composição bentônica e transectos com correntes durante um ano em recifes rochosos de três ilhas (Menina, Mãe e Pai). Um total de 2466 peixes pertencentes a 29 famílias e 42 espécies foi registrado. Os peixes mais abundantes foram *Parablennius pilicornis*, *Haemulon steindachneri*, *Orthopristis ruber* e *Diplodus argenteus*. Os recifes mais abrigados e com maior complexidade de habitats foram os que apresentaram maior abundância e diversidade de espécies. Análises estatísticas revelaram distinções significativas entre as áreas amostradas e um padrão sazonal secundário. As três áreas apresentaram basicamente a mesma comunidade de peixes, estruturadas a partir de gradientes ambientais e de características estruturais próprias de cada ilha.

**Key words:** Rocky reef fishes, Habitat complexity, Benthic biota.

### Introduction

Most studies about the community ecology of Brazilian reef fishes began in the late 1990s, and applied underwater visual census techniques (Rosa & Moura, 1997; Rocha *et al.*, 1998; Ferreira *et al.*, 1998; Rocha & Rosa, 2001). These studies were conducted at several sites throughout the tropical and subtropical Brazilian coastline, from below the mouth of the Amazon River at the Manuel Luiz Reefs (0°52' S) (Rocha & Rosa, 2001), to the south to Santa Catarina State (27°30' S) (Floeter *et al.*, 2001).

On the southeastern Brazilian coastline, rocky shores comprise the habitat for reef fishes and other reef-associated biota (Floeter *et al.*, 2004; Rangel *et al.*, 2007). Despite an overall lower complexity when compared to coral reefs, rocky reefs are characterized by a high fish diversity and endemism (Ferreira *et al.*, 2001; Rocha, 2003). Considering the continental dimension of the Brazilian coast (almost 8000 km), a limited set of scientific works on the ecology and community struc-

ture of reef fishes is currently available.

Differences in reef fish communities can be found along geographic gradients (Parrish, 1989; Sale, 1991; Joyeux *et al.*, 2001) and distinct reef types (Ebeling & Hixon, 1991; Ferreira *et al.*, 2001; Bellwood & Wainwright, 2002). Patterns of occurrence and distribution of these assemblages often vary over space and time, influenced by a set of integrated physical and biological factors (McCormick, 1989; Syms & Jones, 2000; Acosta & Robertson, 2002). Wave exposure, water depth and topographic complexity may exert variable degrees of interference upon many ecological processes, such as recruitment, predation and territoriality (Caley, 1993; Cheney & Côté, 2003; Kawasaki *et al.*, 2003).

In this work, we analyzed reef fish community variation over space and time on three islands in Itaipu Sound, Rio de Janeiro, Brazil. We studied changes in species composition, abundance and diversity and their correlation with benthic coverage and habitat complexity.

## Materials and Methods

### Study Site.

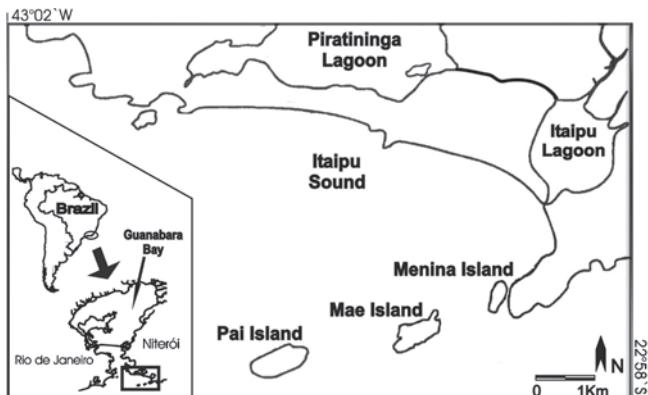
This study was carried out on three islands of Itaipu Sound, Niterói, RJ, Brazil ( $22^{\circ}58'S$ ;  $43^{\circ}02'W$ ) from April 2002 to March 2003. Menina Island, M  e Island and Pai Island are, respectively, 1 km, 2.5 km and 4.2 km off the Brazilian coast. Pai Island is the farthest and most exposed study site. We selected one sampling station located on the sheltered face of each island (Fig. 1). This selection was based on water visibility and diving conditions which could enable surveys to be conducted throughout the year.

### Environmental variables and benthic biota.

At each site, we measured water visibility (Secchi disk), water temperature at diving depth ( $^{\circ}C$ , mercury thermometer) and site rugosity using the chain method (Luckhurst & Luckhurst, 1978) as an index of habitat complexity. Benthic organisms were pooled into ecological and morpho-functional groups: grazers (black and purple sea urchins), sessile constructive (mussels, oyster and barnacles), mobile (gastropoda, sea cucumbers and star fish) and sessile (anemones, turf algae and bryozoans). Individual counts and percent coverage were conducted on replicate 1-m<sup>2</sup> aluminum squares (quadrats) at different points within the same area where visual censuses for reef fishes were performed.

### Reef fish community.

Reef fish community was assessed through 20 m x 2 m underwater strip transects (40 m<sup>2</sup>), varying from the shallowest zones to the bottom rocky/sand interface, about 6-9 meters deep. All sites were visited once a month, and censuses were conducted within the same day between 8:00 and 12:00, starting at Pai Island and finishing at Menina Island. We conducted a total of 108 transects, 36 on each site. Fish counts were conducted in a slow constant swimming velocity (Russ & Gladstone, 1990) and performed by the same observer (J. P. Mendon  a-Neto). Sighted fishes were identified and counted,



**Fig. 1.** Map of Itaipu Sound, showing the three study sites: Menina, M  e and Pai Islands. Inset represents the geographical location of Itaipu Sound in relation to Guanabara Bay and the Rio de Janeiro coast within Brazil.

and the information was recorded on a PVC plate (Williams, 1982). Afterwards, fish count data were expressed as frequency of occurrence and relative abundance.

### Statistical analysis.

Renyi's graphic ordination method was used to check if species diversity was comparable between sites (Hill, 1973; Legendre & Legendre, 1998). According to the method, when diversity patterns are parallel to each other, species diversity may be compared, but when diversity patterns cross each other, no comparisons are possible. Differences among average diversity were statistically tested with bootstrapping following the model proposed by Solow (1993). The Species Diversity & Richness v.2.3 software was used for all procedures (Henderson & Seaby, 1998).

Multivariate analysis of the data including non-metric multidimensional scaling (nMDS) and analysis of similarities (ANOSIM) was performed with PRIMER 5 software (Primer-E 2000). Both procedures used a Bray-Curtis similarity matrix. MDS is a 3-dimensional ordination of samples brought down to a 2-dimensional plot. The quality of the plot is indicated by its stress value: values  $<0.2$  give a potentially useful 2-dimensional picture, stress  $<0.1$  corresponds to a good ordination, and stress  $<0.05$  is an excellent representation. The ANOSIM significance test compares similarities between species by sample matrices. The global R indicates the degree of similarity between the tested groups, with values between -1 and 1. If all replicates within sites are more similar to each other than any replicate from different sites, the value of R is 1; values close to zero indicate that the similarity between sites is very high. A two-way crossed analysis of ANOSIM was performed with the original data, where no transformation or standardization was conducted (Clarke & Warwick, 1994).

The relationship between fish community and environmental variables (water visibility, temperature, bottom complexity and benthic coverage) was investigated by canonical correspondence analysis (CCA) (Legendre & Legendre, 1998). The primary matrix consisted of abundance data of fish species (row) per sample (column), and the secondary matrix environmental consisted of data including the variables previously mentioned and the relative abundance of benthic coverage data. Monte Carlo permutation tests were employed to determine the statistical significance of the main correlations, and their relative contribution to the final model. Only statistically significant variables were maintained. A similar procedure was applied to determine the statistical significance of the first four canonical axes. This analysis was entirely performed on CANOCO for Windows 4.0 (ter Braak & Smilauer, 1998).

## Results

A total of 2446 fishes from 29 families and 42 species were recorded through visual censuses at the three sites during the survey (Tab. 1). *Parablennius pilicornis* was the most abundant species in all sites pooled, followed by *Haemulon*

**Table 1.** Fish families, species, symbols, number of individuals (n) and relative abundance (%) at each site (Pai, Mæ, and Menina Islands) and pooled for all sites (Total).

Family	Species	Symbol	Pai Island		Mæ Island		Menina Island		Total		
			n	%	n	%	n	%	nT	%T	%F
Acanthuridae	<i>Acanthurus chirurgus</i>	ACACHI	5	0.65	11	1.03	8	1.26	24	0.97	22.2
Ariidae	<i>Bagre</i> spp.	BAGSPP	0	0.00	4	0.38	0	0.00	4	0.16	5.6
Belonidae	<i>Hemiramphus</i> spp.	HEMSPP	0	0.00	9	0.85	0	0.00	9	0.36	8.3
Blennidae	<i>Parablennius pilicornis</i>	PARPIL	147	19.09	182	17.12	159	25.12	488	19.79	80.5
Blennidae	<i>Scartella cristata</i>	SCACRI	25	3.25	10	0.94	8	1.26	43	1.74	52.7
Blennidae	<i>Parablennius</i> spp.	PARSPP	6	0.78	3	0.28	1	0.16	10	0.41	13.8
Bothidae	<i>Bothus ocellatus</i>	BOTOCE	0	0.00	11	1.03	0	0.00	11	0.45	16.6
Carangidae	<i>Caranx</i> spp.	CARSPP	1	0.13	2	0.19	30	4.74	33	1.34	8.3
Carangidae	<i>Pseudocaranx dentex</i>	PSEDEN	5	0.65	5	0.47	0	0.00	10	0.41	16.6
Chaetodontidae	<i>Chaetodon striatus</i>	CHASTR	16	2.08	45	4.23	41	6.48	102	4.14	80.5
Chaetodontidae	<i>Chaetodon sedentarius</i>	CHASED	0	0.00	0	0.00	3	0.47	3	0.12	2.8
Dactylopteridae	<i>Dactylopterus volitans</i>	DACVOL	41	5.32	66	6.21	44	6.95	151	6.12	86.1
Diodontidae	<i>Chylomicterus spinosus</i>	CHYSPI	5	0.65	21	1.98	0	0.00	26	1.05	36.1
Diodontidae	<i>Diodon hystrix</i>	DIOHYS	0	0.00	0	0.00	1	0.16	1	0.04	2.8
Gerreidae	<i>Eucinostomus</i> spp.	EUCSPP	7	0.91	9	0.85	0	0.00	16	0.65	13.8
Haemulidae	<i>Haemulon steindachneri</i>	HAESTE	131	17.01	77	7.24	61	9.64	269	10.91	83.3
Haemulidae	<i>Orthopristis ruber</i>	ORTRUB	119	15.45	74	6.96	67	10.58	260	10.54	86.1
Haemulidae	<i>Anisotremus virginicus</i>	ANIVIR	5	0.65	1	0.09	0	0.00	6	0.24	13.8
Holocentridae	<i>Myripristes jacobus</i>	MYRJAC	0	0.00	2	0.19	0	0.00	2	0.08	2.8
Labridae	<i>Halichoeres poeyi</i>	HALPOE	6	0.78	38	3.57	2	0.32	46	1.87	41.6
Labridae	<i>Halichoeres dimidiatus</i>	HALDIM	0	0.00	9	0.85	0	0.00	9	0.36	8.3
Labrisomidae	<i>Labrisomus nuchipinnis</i>	LABNUC	1	0.13	14	1.32	7	1.11	22	0.89	38.8
Labrisomidae	<i>Malacoctenus</i> spp.	MALSPP	0	0.00	0	0.00	4	0.63	4	0.16	2.8
Lutjanidae	<i>Lutjanus</i> spp.	LUTSPP	0	0.00	6	0.56	3	0.47	9	0.36	11.1
Monacanthidae	<i>Stephanolepis hispidus</i>	STEHIS	57	7.40	54	5.08	37	5.85	148	6.00	88.8
Mullidae	<i>Pseudupeneus maculatus</i>	PSEMAC	11	1.43	32	3.01	6	0.95	49	1.99	47.2
Muraenidae	<i>Gymnothorax</i> spp.	GYMSPP	0	0.00	5	0.47	0	0.00	5	0.20	8.3
Pomacanthidae	<i>Pomacanthus paru</i>	POMPAR	0	0.00	4	0.38	1	0.16	5	0.20	11.1
Pomacentridae	<i>Abudefduf saxatilis</i>	ABUSAX	29	3.77	91	8.56	2	0.32	122	4.95	52.7
Pomacentridae	<i>Chromis multilineata</i>	CHRMLUL	38	4.94	26	2.45	3	0.47	67	2.72	41.6
Pomacentridae	<i>Stegastes fuscus</i>	STEFUS	1	0.13	4	0.38	21	3.32	26	1.05	25.0
Priacanthidae	<i>Heteropriacanthus cruentatus</i>	HETCRU	0	0.00	2	0.19	0	0.00	2	0.08	2.8
Scaridae	<i>Sparisoma axillare</i>	SPAAXI	0	0.00	8	0.75	0	0.00	8	0.32	11.1
Scianidae	<i>Pareques aequinatus</i>	PARACU	12	1.56	24	2.26	2	0.32	38	1.54	44.4
Scorpaenidae	<i>Scorpaena</i> spp.	SCOSPP	0	0.00	1	0.09	1	0.16	2	0.08	5.6
Serranidae	<i>Mycteroperca acutirostris</i>	MYCACU	10	1.30	22	2.07	3	0.47	35	1.42	50.0
Serranidae	<i>Epinephelus marginatus</i>	EPIMAR	4	0.52	21	1.98	3	0.47	28	1.14	36.1
Sparidae	<i>Diplodus argenteus</i>	DIPARG	62	8.05	88	8.28	49	7.74	199	8.07	83.3
Sparidae	<i>Pagrus pagrus</i>	PAGPAG	0	0.00	6	0.56	0	0.00	6	0.24	2.8
Synodontidae	<i>Synodus</i> spp.	SYNSPP	0	0.00	4	0.38	1	0.16	5	0.20	11.1
Tetraodontidae	<i>Sphoeroides spengleri</i>	SPHSPE	26	3.38	66	6.21	65	10.27	157	6.37	83.3
Tetraodontidae	<i>Sphoeroides greeleyi</i>	SPHGRE	0	0.00	6	0.56	0	0.00	6	0.24	8.3
Total			770	100	1063	100	633	100	2466	100	

**Table 2.** Summary of fish community (means  $\pm$  standard deviation), habitat complexity and benthic composition (relative percent abundance) plus ANOVA results among sites. \* Significant difference.

	Pai (P)	Mæ (M)	Menina (Me)		
Fish species richness (species per census)	11 $\pm$ 7.7	19 $\pm$ 6.3	10 $\pm$ 4.9	p<0.05*	M≠P=Me
Total number of fishes per census	64 $\pm$ 4	88 $\pm$ 7.6	53 $\pm$ 4.9	p<0.01*	M≠P=Me
Habitat complexity (chain link rugosity index)	1.24 $\pm$ 0.1	1.36 $\pm$ 0.09	1.17 $\pm$ 0.17	p<0.05*	M≠P=Me
GRAZERS	%	%	%		
Black sea urchin - <i>Echinometra lucunter</i>	29.3	9.4	31.2	p<0.05*	M≠P=Me
Purple sea urchin - <i>Litechinus variegatus</i>	15.2	10.9	13.7	p=0.2	M=P=Me
SESSILE CONSTRUCTIVE					
Mussel - <i>Perna perna</i>	3.8	5.6	22.1	p<0.01*	M=P≠Me
Oyster - <i>Crassostrea</i> spp.	2.9	-	13.1	p<0.05*	M≠P≠Me
Barnacle - <i>Balanus</i> spp.	3.1	5.7	5.2	p<0.01*	M=P=Me
MOBILE					
Sea cucumber - <i>Isthicopuss badionatus</i>	0.6	14.3	-	p<0.01*	M≠P=Me
Star fish - <i>Echinaster brasiliensis</i>	3.1	1.8	-	p=0.72	M≠P≠Me
Sessile invertebrates - Gastropoda	2.3	2.9	-	p=0.48	M=P≠Me
SESSILE					
Anemone - <i>Bunodosoma caissarum</i>	9.1	7.3	1.6	p<0.01*	M=P≠Me
Turf algae	30	39.1	12.7	p=0.07	M=P≠Me
Bryozoan - <i>Schizophorella</i> spp.	0.6	3	-	p=0.05*	M≠P=Me

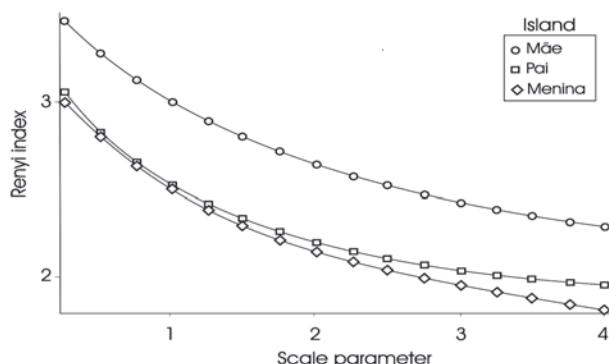
*steindachneri*, *Orthopristis ruber* and *Diplodus argenteus*. Together, they comprised 49% of the total number of fishes recorded. *Stephanolepis hispidus* was the most frequent species at all sites, followed by *Dactylopterus volitans* and *Orthopristis ruber*. The three sites sustained basically the same community of fish species, despite differences in abundance (Table 1). Fish species richness and total number of fishes per census were significantly higher at M  e than at the Pai and Menina sites (Table 2). Reyn  s diversity showed that M  e reached a significantly higher diversity when compared with Menina and Pai Islands (Fig. 2).

A total of 770 fishes belonging to 25 species of 17 families were sighted at Pai Island. The top three species in terms of relative abundance were *Parablennius pilicornis*, *Haemulon steindachneri* and *Orthopristis ruber*. Together they represented 51% of all observations (Table 1). This site was characterized by a mean water visibility of 5.7 m and low habitat complexity; the greatest proportion of the benthic community was represented by grazers and turf algae (Table 2).

At M  e Island, a total of 1063 fishes from 39 species and 27 families were recorded. *Parablennius pilicornis*, *Abudefduf saxatilis* and *Diplodus argenteus* were the most abundant species, totaling 34% of the local abundance (Table 1). This site showed the best water visibility, varying between 6 and 12 m, and the highest habitat complexity. Turf algae and gastropod mollusks had the greatest abundance among members of the benthic community (Table 2).

Menina Island showed intermediate species richness and the lowest overall abundance (633 fishes from 28 species and 21 families). *Parablennius pilicornis*, *Orthopristis ruber* and the *Sphoeroides spengleri* were the most abundant species and accounted for 46% of the total sightings at this site (Tab. 1). Menina Island was characterized by the lowest water visibility, with a mean of 4.6 m, often at the technical limits for visual census, and lowest habitat complexity. Grazers and sessile constructive organisms dominated the benthic community (Table 2).

Multidimensional scaling revealed an apparent pattern of separation between sampling sites, with M  e clearly differing from both Pai and Menina Islands. A secondary seasonal pattern was less evident, suggesting a separation between

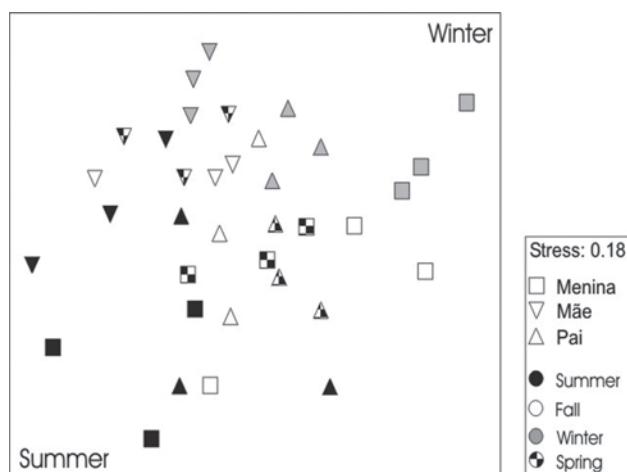


**Fig. 2.** Renyi  s diversity ordination scale parameter comparing Menina, M  e and Pai Islands.

summer and winter samples (Fig. 3).

The two-way analysis of similarities (ANOSIM) further indicated that significant spatial differences were predominant, nevertheless with some degree of overlapping between factor levels. Pairwise tests showed that M  e and Menina Islands were significantly different from each other. Differences between M  e and Pai Islands and Pai and Menina Islands were less evident. Seasonal analysis was also significant, but results were less conclusive due to the high degree of overlapping between factor levels. Pairwise comparisons suggested that summer samples were significantly discernible from a pooled group of fall – spring samples (Table 3).

Results from the CCA analysis yielded a total explained variance of 15.1 % at the first two canonical axes. Significant relationships between species distribution and environmental variables were restricted to four variables including bottom complexity, water temperature, percent cover of grazers and sessile constructive organisms (Fig. 4). The first canonical axis explained 8.5% of the total variance (Monte Carlo,  $p = 0.02$ ) and was positively correlated with the 'x' component of



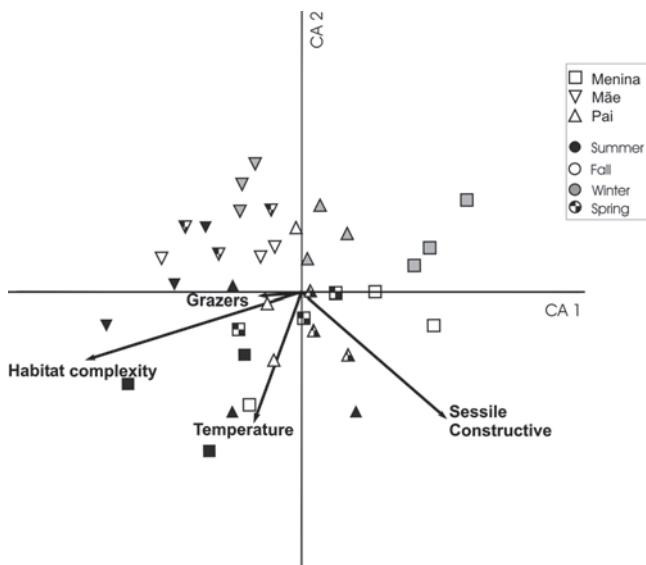
**Fig. 3.** Nonmetric multidimensional scale (nMDS) analysis of individual samples from Menina, M  e and Pai Islands indicating a Summer – Winter seasonal gradient.

**Table 3.** Results of the two-way crossed analysis of similarity considering sites (Menina, M  e and Pai Islands) and seasons (fall, winter, spring and summer) as the main effects.

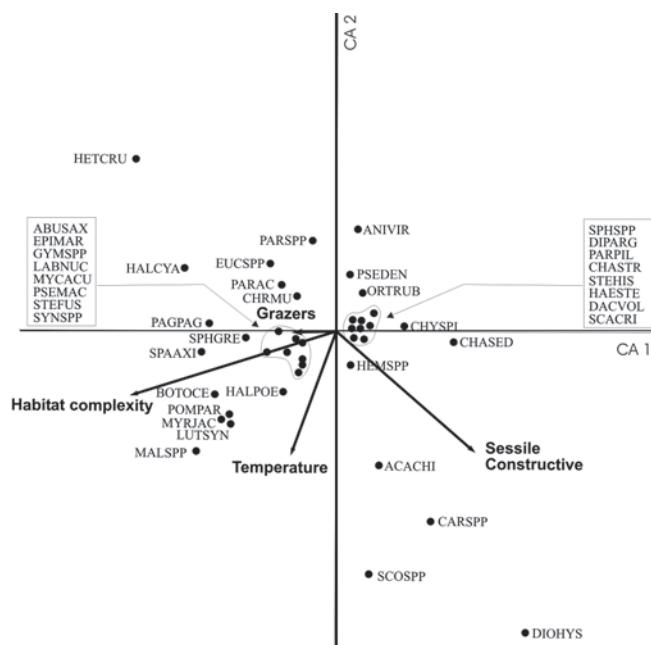
Factors	R Statistic	Significance
Sites		
Global	0.461	0.001
Pairwise Comparisons		
Pai vs M��e	0.407	0.006
Pai vs Menina	0.269	0.011
M��e vs Menina	0.694	0.001
Season		
Global R	0.305	0.001
Pairwise Comparisons		
Fall vs Winter	0.247	0.005
Fall vs Spring	0.049	0.347
Fall vs Summer	0.296	0.058
Winter vs Spring	0.370	0.006
Winter vs Summer	0.580	0.007
Spring vs Summer	0.358	0.012

sessile constructive benthic organisms, and negatively correlated with grazers and the 'x' component of both bottom complexity and water temperature. The second axis explained 6.6% of the total variance (Monte Carlo,  $p = 0.02$ ) and showed a negative correlation with the 'y' projection of all variables (Figs. 4 and 5).

Samples from Mäe Island were mostly separated from the other two sites in the two-dimensional canonical space, were



**Fig. 4.** Canonical Correspondence Analysis (CCA) of individual samples from Menina, Mäe and Pai Islands, indicating a Summer – Winter seasonal gradient and the correlated variables that most probably explain sample distribution in the canonical space.



**Fig. 5.** Canonical Correspondence Analysis (CCA) of species and the correlated variables that most probably explain species distribution in the canonical space.

negatively associated with the presence of sessile constructive organisms, and partly associated with the 'x' component of bottom complexity. Separation between samples from the Pai and Menina Island sites was less evident, but a seasonal gradient was apparent, with a tendency of winter samples to occur on the positive side of the canonical axis 2. Summer samples though followed the water temperature vector in the canonical space (Fig. 4).

Most species clustered in the middle of the diagram indicating a weak association with both principal axes. Nevertheless, *Acanthurus chirurgus*, *Diodon Hystrix*, *Caranx* spp. and *Scorpaena* spp. were associated with spring and summer samples at Pai Island and correlated with the presence of the sessile constructive benthos. *Bothus ocellatus*, *Pomacanthus paru*, *Halichoeres poeyi*, *Myripristis jacobus*, *Lutjanus synagris* and *Malacoctenus* spp. were predominant in summer samples at Menina Island, whereas *Heteropriacanthus cruentatus*, *Parablennius* spp., *Halichoeres dimidiatus*, *Eucinostomus* spp. were mostly associated with Mäe Island samples (Fig. 5).

## Discussion

Underwater censuses at Pai, Mäe and Menina Islands in Itaipu yielded a total of 42 species and 29 families. Our results suggest that our three study sites display low reef fish species diversity when compared to other sites where reef fish studies have been conducted in Brazil (e.g., Floeter *et al.*, 2001). Compared with lower latitude sites, these differences may be explained by the natural north-south latitudinal decline in species richness of reef fishes on the Brazilian coast, a consequence of regional environmental gradients and local processes (Rocha, 2003; Floeter *et al.*, 2004). Also, due to geographic location, the studied site is potentially under the influence of water discharge from the Guanabara Bay (Mendonça Filho *et al.*, 2003; Vilela *et al.*, 2003). The contributions of low-salinity waters, high sediment loads, and pollution runoff derived from such a large body of water, may cause low species richness, as other authors have observed in similar habitats (Roberts & Hawkins, 1999; Silva & Araújo, 2003). In fact, such influences tend to be higher with strongly site-attached fish species, as most of the reef fishes are, than with species having a wider home range (Roberts & Ormond, 1987).

The most abundant species in our study sites were representative of the coastal rocky reef fish fauna of SE Brazil (Ferreira *et al.*, 2001; Ferreira *et al.*, 2004; Rangel *et al.*, 2007). Species such as *Haemulon steindachneri*, *Orthopristis ruber* and *Diplodus argenteus* are frequently captured by local artisanal fishers with hook and line or gill nets near the reefs (Tubino *et al.*, 2007). The blenny, *Parablennius pilicornis*, was the dominant species at all sites, probably reflecting the shallow and sheltered habitat characteristics (Schofield, 2003). Nevertheless, the low density and absence of natural predators, such as groupers and large piscivores, may probably lead to a greater abundance of small blenniid fishes. Large-

sized fishes were nearly absent from the sites, but occurred sporadically in the local artisanal fisheries landings (Tubino *et al.*, 2007), suggesting that fishing pressure reduced their abundance in more accessible and protected sites at the islands. Floeter *et al.* (2006) provided evidence that fishing pressure on target species leads to unbalanced reef fish communities, with a possible reduction of top-down control mechanisms.

There are several possible factors that may determine reef fish distribution on Itaipu islands. Local differences observed among sites reflected spatial patterns and seasonal changes associated with structural complexity and species life histories. Several authors have tried to explain the structure of reef fish communities by looking at their exposure to wave energy (*e. g.* Williams, 1982; Russ, 1984; Galzin, 1987). Although we searched for sheltered sites with comparable hydrodynamics, our results suggested a subtle exposure gradient existed from the most exposed, Pai Island, to the least exposed, Menina Island. Wave exposure in Pai Island may produce a significant difference in the benthic composition, leading to decreased habitat complexity when compared to the other two sites.

CCA analysis further evidenced the influence of habitat complexity upon species abundance and diversity. We observed a distribution pattern with the most abundant, sighted species being grouped closer to the center of the CCA axis (e.g., *Orthopristis ruber* and *Halichoeres poeyi*), while rare, least abundant species were distributed farthest from the CCA center (e.g., *Heteropriacanthus cruentatus* and *Diodon hystrix*). Close to the CCA center, we observed the presence of two major groups, one formed by benthic invertebrate feeders (Haemulidae, *Diplodus argenteus* and *Parablennius pilicornis*), and the other by territorial herbivores (damselfishes, *Stegastes fuscus*) and resident predators (groupers, *Epinephelus marginatus*). Both groups, considered key species, greatly influence reef fish communities (Almany & Webster, 2004). Nevertheless, their ecological role at Itaipu appears less important due to their central distribution in the canonical space. The high fishing pressure upon top predators in the area may explain, at least, the low abundance of groupers, minimizing their role as a key species in the reef fish community at the studied area.

The highest diversity was attained at M  e Island; probably as a consequence of the greatest habitat structural complexity. Habitats with high structural complexity usually support more species and individuals by providing shelter and refuge against predation and competition (Holt, 1987; Greenberg *et al.*, 1995; Beukers & Jones, 1997; Connell, 1997; Charbonnel *et al.*, 2002; Gratwicke & Speight, 2005). Complexity appears to have a marked influence on abundance of territorial and sedentary fish species, such as damselfishes (Pomacentridae) and groupers (Serranidae) (Sale & Douglas, 1984). Also, sheltered habitats may be a bottleneck for recruitment and growth of many reef species (Williams, 1991; Montgomery *et al.*, 2001). The presence of large-sized rocks forming burrows and shelters at different size scales was a

Fish fauna at Menina Island was depauperate compared with the other sites probably due to its proximity to the shore, suggesting that both intensive fishing and coastal runoff influenced fish community (Dulvy *et al.*, 2002). Menina Island was also the most visited site by artisanal fishers and weekend divers (Mendonça-Neto, J. P. *pers. observation*).

Seasonal patterns were observed as a reflection of abundance and occurrence of some species throughout the year, with a clear distinction of species composition during summer and winter months. The observed seasonal tendencies were greater in species diversity during the summer months at all of the sites. This tendency may be related to the cyclic upwelling phenomena which occurs in summer months in the Brazilian southeastern region, bringing cold and nutrient-rich waters close to the coastal zone. Other studies on this region showed that plankton biomass, production and consumption were higher in summer months than in winter (Freitas & Muelbert, 2004).

Our work showed that despite the fact that reef fish community within the Itaipu coastal zone showed similar elements, differences between sites were evident, mostly mediated by differences in the structural patterns of the rocky reefs, but also by environmental factors that vary considerably from one site to another and also throughout the year.

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